Algal Biomass Production via Open Pond Algae Farm Cultivation: 2020 State of Technology and Future Research

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National Renewable Energy Laboratory

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List of Acronyms

AFDW	ash-free dry weight
ANL	Argonne National Laboratory
ARID	Algal Raceway Integrated Design
ASU	Arizona State University
ATP ³	Algal Testbed Public-Private Partnership
AzCATI	Arizona Center for Algae Technology and Innovation
BAT	Biomass Assessment Team
BETO	Bioenergy Technologies Office
CAP	combined algae processing
CO ₂	carbon dioxide
DISCOVR	Development of Integrated Screening, Cultivar Optimization, and Verification Research
FA	Florida Algae (testbed site under ATP ³ consortium)
FAME	fatty acid methyl ester
FFA	free fatty acid
FY	fiscal year
HCSD	high-carbohydrate <i>Scenedesmus</i>
LCA	life cycle assessment
MBSP	minimum biomass selling price
MFSP	minimum fuel selling price
NREL	National Renewable Energy Laboratory
PNNL	Pacific Northwest National Laboratory
SOT	state of technology
TEA	techno-economic analysis

Executive Summary

The annual State of Technology (SOT) assessment is an essential activity for platform research conducted under the Bioenergy Technologies Office (BETO). It allows for the impact of research progress (both directly achieved in-house at NREL and furnished by partner organizations) to be quantified in terms of economic improvements in the overall biofuel production process for a particular biomass processing pathway, whether based on terrestrial or algal biomass feedstocks. As such, initial benchmarks can be established for currently demonstrated performance, and progress can be tracked toward out-year goals to ultimately demonstrate economically viable biofuel technologies.

NREL's algae SOT benchmarking efforts focus both on front-end algal biomass production and separately on back-end conversion to fuels through NREL's "combined algae processing" (CAP) pathway. The production model is based on outdoor long-term cultivation data, enabled by comprehensive algal biomass production trials conducted under the Development of Integrated Screening, Cultivar Optimization, and Verification Research (DISCOVER) consortium efforts, driven by data furnished by Arizona State University (ASU) at the Arizona Center for Algae Technology and Innovation (AzCATI) testbed site. The CAP model is based on experimental efforts conducted under NREL research and development projects.

This report focuses on front-end algal biomass production, documenting the pertinent algal biomass cultivation parameters that were input to the NREL open pond algae farm model. Through partnerships under DISCOVER, collaborators at ASU furnished details on cultivation performance metrics including biomass productivity and harvest densities for recent growth trials done at the AzCATI site. The resulting biomass productivity rates were calculated as 18.4 g/m²/day (ash-free dry weight [AFDW], annual average) for seasonal cultivation of *Scenedesmus obliquus* UTEX393, *Monoraphidium minutum* 26B-AM, and *Picochlorum celeri* biomass strains, grown in September (fall) and April–May (spring), October through March (fall through spring), and June through August (summer), respectively, at the ASU site. After incorporating the production data into a techno-economic analysis (TEA) model for algal biomass production based on a hypothetical commercial facility consisting of 5,000 acres of cultivation pond area (based on NREL's 2016 algae farm design case), the resulting **minimum algal biomass selling price (MBSP) was estimated at \$683/ton** (AFDW basis) in 2016 dollars assuming "*n*th plant" economics for a mature facility utilizing low-cost unlined ponds, coupled with a targeted biomass composition consistent with NREL's high-carbohydrate *Scenedesmus* (HCSD) projections to ensure consistent nutrient costing versus downstream recycle credits from conversion operations. Alternatively, a scenario assuming the use of fully lined ponds would translate to an SOT biomass cost of \$853/ton. Another alternative scenario was also considered based on evaporation rates and salt blowdown disposal requirements reflective of the Algal Testbed Public-Private Partnership (ATP³) consortium's previous Florida Algae (FA) site (the basis for prior 2015–2016 SOT data before being decommissioned and unavailable for later SOTs), which would reduce **MBSP to \$603/ton** for the unlined pond case or \$772/ton for the lined case, given significantly lower net evaporation rates (evaporation minus precipitation) and thus salt accumulation levels in the ponds, a critical factor to consider for saline cultivation.

Relative to the fiscal year (FY) 2019 SOT at \$764/ton or \$670/ton for ASU and FA evaporation scenarios, respectively, (unlined pond basis), the FY 2020 SOT substantially improves on this

result by 11% and 10%, respectively. This is primarily attributed to a **notable 16% improvement in annual cultivation productivity** achieved at the AzCATI site (supported by the efforts under the DISCOVER consortium noted above focused on this metric) through cultivation campaigns aimed at maximizing seasonal strain performance based on known high-performing strains for given seasonal temperatures, as well as pond operational optimizations that were anticipated to support productivity enhancements and mitigate contamination. Coupled with improved operational understanding and logistical support for long-term cultivation trials, this translated to demonstrating such productivity improvements simultaneously alongside an 8% increase in the underlying experimental on-line time, with the number of cultivation production days behind the seasonal productivity data growing from 328 (FY 2019) to 353 (FY 2020), exceeding NREL's n^{th} -plant model basis fixed at 330 days per year of production uptime. After including downstream dewatering/blowdown and short-term storage losses, the overall modeled biomass production output to conversion was calculated at 26.7 tons/acre-yr for both the ASU and FA evaporation basis. A key factor behind the continued improvements to cultivation productivity was likely the use of a fungicide to control contamination and maintain culture health during spring and fall seasons, as well as new introduction of the *Picochlorum* strain optimally suited for hot summer temperatures, achieving exceptionally high growth rates without notable contamination events owing to high salinity conditions. While the cost of fungicide utilized experimentally was not explicitly included in this n^{th} -plant analysis, a sensitivity case estimates that its usage would incur a trivial penalty of roughly \$8/ton to overall MBSPs.

Finally, this milestone reports on key process sustainability indicators for the biomass production stage including annual biomass yields, facility power demand, and water consumption. In keeping with recent BETO guidance, formal life cycle assessment sustainability metrics such as greenhouse gas emissions or fossil energy consumption are not calculated here, but will be deferred to Argonne National Laboratory collaborators.

While significant continued progress has been maintained in supporting high biomass cultivation growth rates, further work remains moving forward to fill in modeling gaps and ultimately demonstrate achievement of future performance targets. In order of importance, these include:

- 1) Increasing focus on achieving higher-quality biomass compositions at harvest (lower protein, higher carbohydrates and lipids) without sacrificing productivity rates
- 2) Incorporating biomass dewatering operations into the scope of SOT data, with further strain prioritization for processability through low-cost dewatering approaches
- 3) Benchmarking current performance for CO₂ retention efficiencies based on system configurations translatable to commercial-scale ponds, and project paths for improvement based on water chemistry, pond operations, or strain/system engineering strategies
- 4) Further focus on high-salinity-tolerant strains, in order to reduce saline water blowdown removal/disposal costs while promoting the use of non-freshwater algae
- 5) Further explore opportunities to optimize harvesting logistics (e.g., batch versus continuous harvesting, harvesting at specific times of day, recycling clarified effluent water from harvest/dewatering operations back to cultivation ponds).

Table of Contents

Introduction	1
Discussion of Relevant Inputs Used in the SOT	3
Results	13
TEA Results	13
Sustainability Metric Indicators	17
Concluding Remarks	18
References	20
Appendix A. TEA Summary Sheet for Base Case Biomass Cultivation SOT Benchmark Model (FA and ASU Evaporation MBSP Scenarios, 2016 Dollars)	22
Appendix B. Life Cycle Inventory for 2020 SOT Algae Farm Model	24

Introduction

The National Renewable Energy Laboratory (NREL) develops and maintains techno-economic models that simulate the technical and economic aspects of conceptual biorefinery conversion pathways to biofuels and bioproducts, focused on both terrestrial and algal biomass processing routes. For a particular set of process parameters, material and energy balance and flow rate information are generated using Aspen Plus simulation software [1] for a given facility size or biomass throughput rate. These data are used to size and cost process equipment and compute raw material and other operating costs. Using a discounted cash flow rate of return analysis, the minimum fuel selling price (MFSP) or minimum biomass selling price (MBSP) required to obtain a net present value (NPV) of zero for a 10% internal rate of return (IRR) is determined. The result is a techno-economic model that reasonably estimates an “ n^{th} -plant” production cost for this pre-commercial process.

Over recent years, NREL has published a number of design reports for both the production of algal biomass and the conversion of algae to fuels via the “combined algae processing” (CAP) pathway [2, 3], both of which focused on out-year targets that, if achieved, would translate to a modeled MBSP of \$494/ton for biomass (2014\$, ash-free dry weight [AFDW] basis) and MFSP of \$5.90 per gallon gasoline equivalent (GGE) for resulting fuels (after revising the original CAP design case to match up with the outputs from the newer algae farm design case, as documented in the 2016 *Multi-Year Program Plan* [4]). The latter MFSP projection was based on NREL’s original CAP processing approach focused on fuels via well-understood conversion technologies, which is evolving toward a focus on hydrocarbon fuels and value-added coproducts to reduce the MFSP toward future targets. However, in order to achieve such fuel cost goals in the future, substantial improvements are required, particularly around biomass cultivation costs, representing the largest contributor to overall fuel cost, driven most strongly in turn by the achievable annual cultivation productivity. Accordingly, this has been the primary parameter of focus in prior algae farm SOTs since 2015, as well as more broadly in the BETO Algae Platform as the subject of numerous funding grants over that time frame.

Upon initiation of algal MBSP benchmarking with the fiscal year (FY) 2015 SOT, the demonstrated annual productivity was 8.5 g/m²/day based on the first year of data generated under a prior consortium titled the Algal Testbed Public-Private Partnership (ATP³), translating to a modeled MBSP of \$1,142/ton in 2016\$. Relative to final future targets of \$488/ton at 25 g/m²/day (updated here to 2016\$ and 21% taxes, versus \$494/ton in 2014\$ noted earlier), this implied a need to improve productivity by roughly threefold in order to reduce MBSP by 60%. Initially, subsequent improvements made after the FY 2015 SOT point were modest relative to the degree of improvement ultimately required, but this in part reflected the fact that the initial focus of ATP³ was strictly to maintain uniformity across testbed sites in establishing transparent benchmarks more than to improve performance. More recently, efforts have shifted to specifically focus on improving cultivation productivity based on hypothesis-driven research to evaluate the most promising strains and cultivation conditions, translating to a more notable improvement in recent SOTs.

The biomass production SOT inputs for the present exercise were all sourced from the Arizona Center for Algae Technology and Innovation (AzCATI) testbed site operating outdoor ponds

over seasonal periods spanning the course of a year. All cultivation trials selected for incorporation in this year's 2020 SOT benchmark leverage Arizona State University's (ASU's) expertise in performing the cultivation work under the support of the Development of Integrated Screening, Cultivar Optimization, and Verification Research (DISCOVER) consortium (<https://discover.labworks.org/>). In keeping with prior SOTs, the cultivation practices and data generation were all based on consistent methods that have been well established by ASU across all season/strain cases. Beyond the crucial cultivation operation itself, other steps in the algae farm model are considered either outside the scope of battery limits (such as CO₂ and nutrient delivery logistics), or otherwise outside the scope of experimental work and therefore available data to which we have access (namely algal biomass dewatering, which was maintained fixed in the biomass production SOT model consistent with the design case). The model will be improved in out-years with the incorporation of relevant data in these areas, replacing the assumptions currently in place.

We emphasize that the present SOT analysis and the resultant MBSP and MFSP values carry some uncertainty related to the assumptions and estimates made for capital and raw material costs. Without a detailed understanding of the underlying basis, the absolute computed selling price has limited relevance. By demonstrating the cost impact of various process parameters individually or in concert, the model helps guide research by indicating where the largest opportunities for cost reduction exist. It is also acknowledged that "state of technology" is arguably a misnomer since no commercial algal biofuel facility exists today (e.g., growing algal biomass for purposes of producing fuels at commercial scale), and because the SOT performance results documented here are based solely on NREL and partner (DISCOVER) data and do not necessarily represent a broader picture of all performers within and beyond BETO's portfolio.

Discussion of Relevant Inputs Used in the SOT

The algal biomass modeling work conducted for this SOT milestone makes use of the prior Aspen modeling framework that was originally established for the 2016 algae farm design report [2, 3]. The process models remain separated between front-end cultivation and dewatering of algal biomass and back-end conversion of biomass via CAP. However, by utilizing the same biomass flow rates, concentrations, and costs (MBSP), as well as pertinent credits for nutrient and CO₂ recycles consistently between the two sides of the process, the resulting MFSP is consistent with a single fully integrated production and conversion facility.

The process schematic for the algal biomass production process as the subject of this SOT discussion is depicted in Figure 1. In summary, the overarching process for the production facility consists of 5,000 acres of production ponds (10 acres each in size) with a total facility footprint of 7,615 acres, coupled to an inoculum propagation system consisting of a series of closed and open growth systems of increasingly larger size, as well as dewatering operations made up of in-ground gravity settlers, hollow fiber membranes, and centrifugation in sequence to ultimately concentrate the biomass from the harvested density up to 20 wt% solids AFDW. The production facility also includes costs for CO₂ (sourced from off-site flue gas carbon capture technology), fertilizer nutrients, delivery pipelines for makeup water from a nearby groundwater resource, and pipelines for on-site culture circulation and CO₂ delivery to ponds. With the ponds representing the critical and most costly step of the process, the n^{th} -plant commercial facility stipulates the use of 10-acre ponds, which are considerably larger than today's "large-scale" standards of 2–3-acre ponds, in order to maximize economy-of-scale benefits, with an additional stipulation that the ponds are unlined (making use of native clay soils) except for small portions of the pond where a plastic liner is used to control erosion. While such a low-cost pond design may reasonably be viewed as representative of a future n^{th} -plant facility, a second alternative scenario also considers the use of fully lined ponds that are more typical in today's early demonstration facilities (or which otherwise may more likely be required in the case of saline cultures). The cost and circulation power demands for the 10-acre ponds are based on average values attributed to four separate pond design estimates that were furnished to NREL from external consultants in support of the 2016 design report.

Figure 1. Schematic diagram summarizing key operations for algae biomass farm process model.

Experimental data outputs to SOT model are primarily focused on the main production pond step, with other operations either considered outside battery limits (CO₂, nutrient, water logistics) or otherwise outside the scope of currently available data (dewatering), and thus set consistent with future design case targets.

As noted above, the inputs for the biomass production model were based on seasonal performance data generated under cultivation trials at the AzCATI testbed site over the past year (data from fall 2019 through summer 2020 feeding the FY 2020 SOT), with the key parameters utilized in the TEA model being productivity rates, biomass density at harvest, and average daily pond evaporation. Additionally, biomass composition estimates are provided here; however, similar to previous SOT practices, the measured composition is based on biomass cultivated under nutrient-replete conditions, translating to high levels of protein and ash but relatively low levels of carbohydrates and lipids. More details on estimated and measured composition are provided below, but in summary, combined carbohydrate and lipid levels generally remain below 30% for the seasonal strains reflected in the FY 2020 SOT, which is impractical for NREL's CAP model in its historical configuration focused to date on these two constituents while protein has traditionally been relegated to anaerobic digestion (AD). Thus, also similar to prior SOTs, the base case FY 2020 SOT model assumes the composition of high-carbohydrate *Scenedesmus* (HCSD) for both the cultivation and CAP conversion process models, given that *Scenedesmus* was the basis used for CAP conversion experiments, and this composition is also consistent with the targeted 2030 goals as described in the algae farm design case [3]. *The SOT baseline cultivation model therefore assumes seasonal productivities, harvest densities, and evaporation rates attributed to the provided cultivation measurements across three seasonally rotated strains (Monoraphidium, Scenedesmus, and Picochlorum), overlaid with HCSD compositional assumptions for nutrient costing. The SOT CAP conversion model assumes this same HCSD composition and resulting seasonal flows from the farm model, with yields across each conversion step set based on experimental data also generally utilizing Scenedesmus.* NREL has recently begun to investigate new CAP configuration options which may allow more feedstock flexibility, including the capability to more effectively utilize higher-protein algae strains [5], but further work and optimization is ongoing.

Details on cultivation protocols and methods, as well as productivity calculations used to inform the SOTs, are consistent with prior SOTs [6–7] and based on work performed by the same partners at ASU. The cultivation experiments are carried out in 4.2-m² open ponds with online monitoring of culture health. Operational conditions include semi-continuous operation over all seasons with up to three times per week harvesting and dilution of the cultures, from which the productivity is calculated as harvest yields based on ash-free dry weight. Additionally, while there was a substantial amount of other experimental activities and strains evaluated under the support of DISCOVER as well as other collaborations making use of ASU's testbed facilities, this milestone report is not intended to provide an exhaustive summary of all such activities. We defer to the associated reports for those respective efforts to provide a more thorough documentation of all activities, methods, hypotheses investigated, lessons learned on what worked and did not work, etc. Only those details as pertinent to the cases/data sets selected to form the basis for the SOT inputs are discussed here.

In summary, the prior warm-season strain used in the 2019 summer rotation (*Scenedesmus obliquus* UTEX393) was maintained for the first month of the fall FY 2020 SOT campaign, with a resulting month-average productivity of 19.4 g/m²/day based on semi-continuous harvesting maintaining a 20-cm pond depth (marking a substantial improvement over September 2018 at 11.0 g/m²/day with *Desmodesmus intermedius* C046 as used for the FY 2019 SOT data set). After that point, with the typical drop-off in seasonal temperatures and solar irradiance observed at the AzCATI site, the site shifted to use of the same cold-season strain *Monoraphidium*

minutum 26B-AM as utilized previously, also reflecting continued improvements over that strain's productivity performance during October and November. This strain was maintained as the "official" SOT basis throughout the winter and into the first month of spring (March 2020), continuing to reflect the best performance observed over those months in colder temperatures and lower irradiance. Similar to cold-season operational practices in prior SOTs, the pond depth was reduced from 20 cm to 10 cm between January and March, running more shallow to increase peak daytime culture temperatures and improve light utilization given lower wintertime solar irradiance. As noted in prior SOT reports, while such shallower depths would not likely be practical in a paddlewheel raceway pond, they could be achieved in sloped gravity-driven pond designs (e.g., the Algal Raceway Integrated Design [ARID] pond system evaluated previously [15]), which have been shown to incur similar overall system costs as paddlewheel raceway ponds as reflected in the 2016 algae farm design report [3]. Moving forward, as experimental data have not conclusively demonstrated notably higher productivity rates for 10 cm versus 20 cm operation during winter months (i.e., limited data comparing 10- vs. 20-cm depths suggested comparable performance during winter 2019 runs and only minimal improvements during winter 2020 runs), moving to shallower winter depths may not be warranted in future SOT campaigns, which would further improve the translatability of the full data set to the commercial n^{th} -plant TEA models utilized here.

Beginning in April as temperatures again warmed, the AzCATI site transitioned back to the UTEX393 strain, utilized as the "official" SOT basis for the remainder of the spring through May given its typically higher performance observed beginning at that point relative to 26B-AM (also maintained at the standard 20-cm pond depth). However, moving into the summer, this strain suffered from substantial contamination events, including from additional contaminants beyond the typical fungal contamination that had previously been successfully controlled with a fungicide (fluazinam) in the FY 2019 SOT campaigns, but which was less successful this time. This would have led to an overall increase in annual downtime and decrease in summer productivity relative to the FY 2019 SOT, but was salvaged by the experimental investigation of a second summer strain, *Picochlorum celeri*. This strain was run over the summer from June through August under high salinity (up to 50 ppt, versus 5-ppt salinity levels run for the UTEX393 and 26BAM strains during other seasonal trials), and demonstrated very high summer growth rates with minimal contamination.

Resulting monthly and season-average productivities over all months utilizing the UTEX393, 26B-AM, and *P. celeri* strains are reported in Table 1. As expected, during the short period with shallower pond operation, harvest densities generally increased over 20-cm operation, translating to somewhat lower downstream processing throughputs through dewatering (however, the dewatering capital expenses were set based on maximum seasonal capacity in the summer and thus this leads to minimal overall MBSP savings). While no fungicide was required with *P. celeri* during the summer months given relatively low contamination pressures (likely in part due to high salinity conditions for this strain), fluazinam was still utilized in the fall, spring, and winter seasons to help promote general culture health, at a 1-ppm dosage approximately every 1.5 weeks in the spring and fall and every 2 weeks in the winter. No contamination events occurred for the seasonal strain selections utilized in this year's SOT data set. Owing to the n^{th} -plant assumptions underpinning the SOT analysis, which in this case makes the assumption that sufficient strain development has occurred to make use of robust algal strains resistant to culture crashes under an n^{th} -plant level of technology maturity, the associated usage and cost for

fungicide is not currently reflected in the present SOT exercise, but related work done under the DISCOVER project’s “operational baseline” assessment found the use of fluazinam at the seasonal dosage levels reflected here would only increase resulting MBSP by \$8/ton [8].

Table 1. Monthly Cultivation Performance for FY 2020 SOT Trials (Source: John McGowen, ASU)

Season	Month	Productivity, g/m ² -day	AFDW at Harvest, g/L	Strain	Days	Season Avg.
Fall	September	19.4	0.35	UTEX393	29	15.0
	October	14.4	0.35	26BAM	30	
	November	11.2	0.36	26BAM	28	
Winter	December	7.1	0.33	26BAM	34	8.3
	January	8.5	0.72 (10 cm)	26BAM	31	
	February	9.4	0.60 (10 cm)	26BAM	28	
Spring	March	14.1	0.63 (10 cm)	26BAM	31	18.5
	April	17.1	0.33	UTEX393	32	
	May	24.2	0.33	UTEX393	30	
Summer	June	27.1	0.40	<i>P. celeri</i>	19	31.6
	July	31.9	0.53	<i>P. celeri</i>	30	
	August	35.8	0.55	<i>P. celeri</i>	31	

Text overlays by harvest density values indicate cold-season months with 10-cm pond depth (otherwise 20-cm pond depth standard). Strain IDs = *Scenedesmus obliquus* UTEX393, *Monoraphidium minutum* 26B-AM, and *Picochlorum celeri*.

Table 2 presents a summary of the cultivation productivity, harvest density, and daily evaporation rates on a seasonal-average basis attributed to the ASU data used in the 2020 SOT, in comparison to prior data used in the 2017–2019 SOTs (generally utilizing 2016–2018 data, respectively) also at the ASU site, as well as the 2015–2016 SOTs (utilizing 2014 and 2015 data, respectively). As noted in prior SOT milestone reports, the first two years constituting the 2015–2016 SOTs were based on cultivation work done at ATP³’s Florida Algae (FA) testbed site, given improved productivities and climate conditions that had been observed at that site while it was operating; however, that site was subsequently decommissioned as the land it occupied was no longer available for ATP³ use, which prompted a change to the ASU testbed site for all cultivation work supporting the 2017 SOT onward. This incurs an obvious but unavoidable disconnect in consistently comparing cultivation performance throughout the full span of the reported years, given different weather variables (solar irradiance, temperatures, seasonal swings, etc.) between the two testbed locations. Accordingly, the SOT continues the prior practice of evaluating costs for the AzCATI-demonstrated productivities overlaid with both ASU and FA seasonal evaporation rates.

Based on the selected cases for the FY 2020 SOT as shown in Table 2, the resulting year-average productivity is 18.4 g/m²/day, which represents a notable 16% improvement over the FY 2019 SOT basis of 15.9 g/m²/day, with improvements in productivity demonstrated for each individual

season aside from a slight drop in spring. Namely, fall and winter season average productivity increased by 32% and 29%, respectively, while spring decreased by 1%; additional further improvements were also made in the summer average productivity, which increased by 16% relative to the FY 2019 SOT to 31.6 g/m²/day, relative to an already significant gain in the summer made during the FY 2019 SOT at 27.1 g/m²/day, driven by the switch to *P. celeris* with ideally suited conditions to this strain favoring high temperatures as exhibited during summer months at the AzCATI site. It is also worth pointing out that these seasonal productivity data were tied to productive use of the ponds nearly 100% of every month except June (and then only because of a late start in beginning the *P. celeris* trials), equating to 353 days of productive cultivation up-time over the full course of FY 2020 (exceeding the fixed n^{th} -plant model assumption at 330 days/year). This latest performance level generally exceeds the best data previously reported elsewhere publicly [9–11], and is based on transparent data and calculation methods provided firsthand. A direct comparison against such other reported values is obfuscated by different locations, pond designs, harvesting protocols, and calculation methodologies for productivity. Also notably, relative to the initial 2015 SOT benchmark at 8.5 g/m²/day, the 2020 SOT case represents over a doubling in this key parameter over five years.

Table 2. Cultivation Productivity (AFDW), Harvest Density (AFDW), and Daily Evaporation Rate for Selected 2020 Cultivation Trials at ASU Site, Compared Against Prior Cultivation Trials at ASU and Florida Algae Sites as well as Alternate FY 2016 SOT Data Furnished by an ABY1 Performer

	Productivity, g/m ² /day	Harvest Density, g/L	Evaporation Rate, cm/day	Algae Strain	Harvests per Week	Harvest Volume, Fraction of Pond	Daily Dilution Rate, Fraction of Pond
2015 SOT (Florida Algae/ATP³)							
Fall 2014	6.8	0.22	0.01	<i>Nanno</i>	1x	0.75	0.11
Winter 2014	5.0	0.23	0.01	<i>Nanno</i>	1x	0.75	0.11
Spring 2014	11.4	0.36	0.14	<i>Nanno</i>	1x	0.75	0.11
Summer 2014	10.9	0.25	0.02	<i>Nanno</i>	1x	0.75	0.11
Average	8.5	0.27	0.04				
2016 SOT (Florida Algae/ATP³)							
Fall 2015	7.0	0.20	0.01	<i>Desmo</i>	3x	0.50	0.21
Winter 2014 ^a	5.0	0.23	0.01	<i>Nanno</i>	1x	0.75	0.11
Spring 2015	11.1	0.28	0.14	<i>Nanno</i>	3x	0.25	0.11
Summer 2015	13.3	0.32	0.02	<i>Desmo</i>	3x	0.50	0.21
Average	9.1	0.26	0.04				
2016 SOT Reference Case (ABY1 Performer)							
Fall	7.8	0.20	0.01	<i>Nanno</i>		Not provided	
Winter	4.8	0.23	0.01	<i>Nanno</i>		Not provided	
Spring	13.0	0.28	0.14	<i>Nanno</i>		Not provided	
Summer	17.5	0.32	0.02	<i>Nanno</i>		Not provided	
Average	10.8	0.26^b	0.04^b				
2017 SOT (ASU/ATP³)							
Fall 2016	8.5	0.30	0.7	<i>Nanno</i>	N/A (batch, harvested every 1–3 weeks)		
Winter 2016	5.5	0.36	0.2	<i>Kirch</i>	N/A (batch, harvested every 2–3 weeks)		
Spring 2016	13.2 (ARID)	0.74	0.9	<i>Scened</i>	5x	0.25	0.18
Summer 2015 ^c	14.1	0.32	1.2	<i>Desmo</i>	3x	0.50	0.21
Average	10.3	0.43	0.7^e				
2018 SOT (ASU/ATP³-DISCOVER-RACER)							
Fall 2016 ^d	8.5	0.30	0.7	<i>Nanno</i>	N/A (batch, harvested every 1–3 weeks)		
Winter 2018	7.7	0.69	0.2	<i>Scened/Monor</i>	N/A (batch, harvested every 10–13 days)		
Spring 2018	15.2	0.70	0.9	<i>Monor</i>	1–3x	0.83	0.17
Summer 2018	15.4	0.35	1.2	<i>Desmo X2</i>	3x	0.55	0.20
Average	11.7	0.51	0.7^e				
2019 SOT (ASU/DISCOVER)							
Fall 2018	11.4	0.41	0.7	<i>Desmo/Monor</i>	2.4x (avg)	0.50	0.17
Winter 2019	6.5	0.51	0.2	<i>Monor</i>	1.3x (avg)	0.65	0.12
Spring 2019	18.7	0.60	0.9	<i>Scened/Monor</i>	2.0x (avg)	0.63	0.18
Summer 2019	27.1	0.43	1.2	<i>Scened</i>	3.0x (avg)	0.75	0.32
Average	15.9	0.49	0.7^e				
2020 SOT (ASU/DISCOVER)							
Fall 2018	15.0	0.35	0.7	<i>Scened/Monor</i>	2.3x (avg)	0.64	0.22
Winter 2019	8.3	0.55	0.2	<i>Monor</i>	1.2x (avg)	0.70	0.12
Spring 2019	18.5	0.43	0.9	<i>Scened/Monor</i>	2.4x (avg)	0.78	0.28
Summer 2019	31.6	0.50	1.2	<i>Pico</i>	3.0x (avg)	0.79	0.34
Average	18.4	0.46	0.7^e				

^a No new winter 2015 data available; winter 2014 data at Florida Algae is maintained for 2016 SOT

^b Harvest densities and evaporation rates were not provided; set consistent with the 2016 ATP³ SOT basis

^c No new summer 2016 data available; summer 2015 data at ASU is maintained for 2017 SOT

^d No new fall 2017 data available; fall 2016 data at ASU is maintained for 2018 SOT

^e Evaporation rate set based on 2017 algae harmonization report for site nearby Phoenix, AZ (30-year average)

The compositional data corresponding with the FY 2020 SOT productivity data shown in Tables 1 and 2 for the DISCOVER cultivation experiments are presented in Table 3. Fall data include the month of September 2019 for *Scenedesmus* UTEX393 (shown in parentheses for the fall column) and October–November 2019 for *Monoraphidium* 26B-AM; compositional data for the fall 26-BAM biomass were not available, so the same data for this strain from the spring were applied given minimal variations for a given strain between seasons. Winter reflects exclusive use of *Monoraphidium* 26B-AM, as well as one month of spring (March 2020, shown in parentheses for the spring column), after which point *Scenedesmus* UTEX393 is reflected for the final two months of spring (April–May 2020). Finally, summer data reflect new use of *Picochlorum celeri* grown in saline conditions. The compositional data shown in Table 3 represent measured averages of 185 samples analyzed. Generally, no significant trends in biomass composition across the four seasons can be observed because of the wide range of physiology underlying the harvested biomass samples. The exception in compositional trends can be found in the total C content of *Monoraphidium* 26B-AM, which is generally higher than the other two strains. Additionally, *P. celeri* exhibits higher ash due to high-salinity conditions for this strain, offset by a reduction primarily in carbohydrate content relative to the other two strains. Further details and caveats behind the compositional values reported here are shown in the footnotes of Table 3.

Similar to previous SOTs, all cultivation experiments were conducted in nutrient-replete conditions, generally translating to similar component compositions as have been reported previously with respect to high protein and relatively low lipid and carbohydrate contents. Namely, fatty acid methyl ester (FAME) lipid content (reported here as free fatty acid [FFA]) is consistently shown to vary between 9 and 10 dry wt%, and fermentable carbohydrates (glucose plus mannose) are between 9 and 12 dry wt% (4% in the case of *P. celeri*), while protein content is higher between 43–48 dry wt%. Although such measured biomass compositions were expected given the focus on maximizing productivity supported by nutrient-replete conditions, they continue to reflect a difference relative to the future target composition exemplified by mid-harvest HCSD projected to be achieved by 2030 (Table 4) with a lower protein content (13 dry wt%) and higher lipid (26% FAME as FFA) and carbohydrate (48%) content [3] (the HCSD composition shown in Table 4 is consistent with the algae farm design report [3], but with additional detail added now to reflect components not previously specified explicitly, such as glycerol and sterols [12]). As the latter for mid- to late-harvest (and thus nutrient-deplete) *Scenedesmus* continues to be the SOT basis for experimental CAP conversion processes given that the experimentally cultivated high-protein material is not practical for such processes (at least based on historical CAP approaches taken to date), as noted above and similar to prior SOT practices, the base case SOT biomass model conducted here maintains the values for seasonal cultivation productivity performance and pond densities as demonstrated at ASU, but overlaid with the HCSD biomass compositional attributes for purposes of running the same HCSD composition through the CAP model as well, and to ensure consistent treatment between raw cultivation nutrient/CO₂ costs versus recycle credits from downstream conversion. The HCSD composition is also consistent with the basis utilized in the 2016 algae farm design case. As an alternate sensitivity case, if the harvested compositions as shown in Table 3 were reflected through the SOT models, the resultant MBSPs would increase by approximately \$104/ton relative to the HCSD basis, primarily by way of increased N/P nutrient demands (although noting that the majority of this increase would subsequently be offset by nutrient recycle credits taken in downstream conversion models).

Table 3. Elemental and Component Compositions Based on Seasonal Average Values for Harvested Strains Reported in Table 1 (Adjusted to 100% Mass Balance Closure) [12, 3]

Elemental (AFDW) ^{a,b}	Fall <i>Scened/Monor</i> ^f	Winter <i>Monor</i>	Spring <i>Monor/Scened</i> ^g	Summer <i>Pico</i>
C	53.2 (52.5)	53.1	52.2 (53.2)	52.3
H	7.8 (7.9)	7.9	7.7 (7.8)	7.5
O	27.4 (27.9)	27.8	27.6 (27.4)	27.8
N	10.2 (10.3)	9.8	11.1 (10.2)	10.9
S	0.2 (0.2)	0.2	0.2 (0.2)	0.2
P	1.2 (1.2)	1.2	1.2 (1.2)	1.2
Total	100.0%	100.0%	100.0%	100.0%
Component (dry wt) ^b				
Ash	6.4 (9.9)	6.5	9.2 (6.4)	18.2
Protein	45.5 (44.3)	43.9	48.0 (45.5)	42.7
FAME lipids ^c	8.6 (9.0)	9.8	9.0 (8.6)	9.1
Glycerol ^c	1.0 (1.0)	1.1	1.0 (1.0)	1.0
Non-fuel polar lipid impurities	5.2 (5.4)	5.9	5.4 (5.2)	5.5
Sterols ^d	0.5 (0.5)	0.5	0.5 (0.5)	0.5
Fermentable carbohydrates ^e	11.3 (11.5)	10.2	9.1 (11.3)	4.3
Other carbohydrates (galactose)	2.3 (2.3)	2.0	1.8 (2.3)	0.9
Cell mass	19.2 (16.0)	20.1	15.9 (19.2)	17.8
Total	100.0%	100.0%	100.0%	100.0%

^a CHN composition is reported as measured CHN data corrected for ash content of the biomass; O was calculated as the difference from mass balance after estimating S and P (as 0.2 and 1.2%, respectively) and adjusted to 100%

^b SOT biomass compositions are less detailed than the HCSD basis; CHN and composition data for *Scenedesmus*, *Monoraphidium*, and *Picochlorum* cases are based on measured averaged data for a total of 185 harvested production samples and are considered representative for primarily nutrient-replete growth conditions. Composition data are currently broken down to ash, protein, lipids as FAME (in this case TAG lipids measured as FAME, with an added estimate of 10% glycerol relative to measured FAME, and an assumed polar lipid headgroup fraction that increased the FAME content by at least 60%), and total carbohydrate content (reported here as 100% fermentable from the measured sum of monosaccharides detected, with an additional 20% non-fermentable and possibly unhydrolyzable or recalcitrant carbohydrates); a remaining component called “cell mass” accounts for between 16% and 19% of the biomass and reflects unidentified components that are not measured but are need to account for the mass balance.

^c Lipids originally characterized as triglycerides (1:1 FAME equivalent); adjusted here to FFA plus glycerol (as ~11% of the measured FAME content, and reflective of actual components in pretreated hydrolysate for *Scenedesmus* biomass.

^d Sterols originally included in “polar lipid impurity” fraction in prior models. Value currently estimated for HCSD, based on a representative earlier-harvest biomass sample; for SOT biomass, sterol concentration is estimated at a flat 0.5% of the biomass, consistent with earlier observations at NREL

^e “Fermentable carbohydrates” typically consist of 75% glucose and 25% mannose for all species analyzed

^f Fall: first value: *Monoraphidium* 26BAM (Oct.–Nov.); value in parentheses: *Scenedesmus* UTEX393 (Sept.); fall data for 26BAM were not available; values above reflect 26BAM data during spring cultivation

^g Spring: first value: *Scenedesmus* UTEX393 (April–May); value in parentheses: *Monoraphidium* 26BAM (March)

Table 4. Elemental and Component Compositions for High-Carbohydrate *Scenedesmus* (HCSD) Biomass (Used for the SOT Base Case Model), Adjusted to 100% Mass Balance Closure, per NREL Algae Farm Design Case [3]

Elemental (AFDW)	HCSD Basis Composition
C	54.0
H	8.2
O	35.5
N	1.8
S	0.2
P	0.22
Total	100.0%
Component (dry wt)	
Ash	2.4
Protein	13.2
FAME lipids	26.0 ^a
Glycerol	3.0 ^a
Non-fuel polar lipid impurities	1.0
Sterols	1.8 ^b
Fermentable carbohydrates	47.8 ^c
Other carbohydrates (galactose)	3.2
Cell mass	1.6
Total	100.0%

^a Lipids originally characterized as triglycerides (1:1 FAME equivalent); adjusted here to FFA plus glycerol (reflective of actual components in pretreated hydrolysate for *Scenedesmus* biomass)

^b Sterols originally included in “polar lipid impurity” fraction in prior models. Value currently estimated for HCSD, based on a representative earlier-harvest biomass sample.

^c “Fermentable carbohydrates” consists of 75.1% glucose, 24.9% mannose

For modeling purposes, the SOT cultivation data for the parameters noted above were input into the “Area 100” section of the biomass production model (cultivation ponds); all other portions of the model were unchanged relative to details described in the design report [3], including makeup CO₂ and water delivery costs to the facility, as well as dewatering design and performance (maintaining the use of in-ground gravity settlers, followed by hollow fiber membranes, and then centrifugation to concentrate the biomass to 1%, 13%, and then 20% AFDW, respectively). In particular, the use of the new *Picochlorum* strain in the FY 2020 SOT data set may incur challenges in dewatering through primary settling, based on qualitative observations for this strain’s settling propensity during the summer cultivation campaign; however, as dewatering remains outside the scope of SOT experimental focus, the basis dewatering approach was left unchanged here (though dewatering operations and/or prioritizing for strains with good settling ability is identified as a key issue for future investigation).

The inoculum system capital and operating costs were maintained at the same fraction of production pond costs as the design case basis. Facility circulation pipelines were resized to reduce pipeline diameters associated with lower overall flows and circulation rates for the SOT models relative to the design case. Additionally, CO₂ utilization in the pond was maintained at an assumed 90% of the feed CO₂. The production ponds assumed in the model were based on 10-

acre individual open raceway ponds, grouped into 50 “modules” within the overall 5,000-acre farm (based on cultivation area).

As noted above, initial SOTs in FY 2015–2016 had utilized cultivation data from the FA testbed site before transitioning to the ASU site for the FY 2017–2020 SOTs, given advantages for the FA site being located in the region (Gulf Coast) that has historically been viewed as most optimal for siting commercial algae farms given high productivities and low water consumption [13, 14]. In addition to the disconnects this switch incurs with respect to locational variables that influence seasonal cultivation productivity, another artifact of the transition to the ASU site which also artificially influences biomass costs is the evaporation rates, which are significantly higher in Arizona than in Florida (where “evaporation rate” here is defined as net evaporation minus precipitation to replenish pond water levels). Namely, the net annual average pan evaporation estimated for the ASU site is 0.73 cm/day, versus 0.04 cm/day previously utilized for the FA site (both largely based on evaporation rates taken from local resource assessment models for each location, again based on net evaporation less precipitation). For saline cultivation as is currently stipulated by BETO to be required for SOTs and design cases moving forward, higher evaporation rates translate to higher blowdown requirements from the system in order to maintain pond salt tolerance limits of the strain. This saline blowdown must be disposed of and cannot merely be discharged to local water bodies, unless the site is located on the coast and can be discharged to the ocean. The current farm models assume the use of deep-well saltwater injection similar to practices employed for hydraulic fracturing in petroleum extraction. At an assumed makeup salt content of 7.7 ppt for locally sourced saline groundwater and an operating expense of \$1.80/m³ blowdown water disposal, *the blowdown requirements for a farm located in Arizona add significant costs to the overall MBSPs relative to a farm located in Florida for saline cultivation scenarios.*

To mitigate this cost as much as possible, the practice is maintained similar to prior SOTs in first routing the blowdown to evaporation ponds to reduce the overall volume of water being disposed of (based on the same seasonal evaporation rates as the production ponds), costed at \$49,455/acre assuming fully lined but simple shallow pits. The ponds are sized to reduce overall water content by 75% (near solubility limits for the dissolved salts). Additionally, the organism salt tolerance was assumed fixed at 50 ppt, which is higher than typical saline strains but within limits recently observed for a hypersaline strain up to 78 ppt and consistent with the salinity levels employed for the *P. celeris* strain in the summer. A second scenario is also considered based on evaporation rates previously modeled for the Florida Algae site, to control for the influence of this variable in the overall MBSP estimates in comparing to the FA basis in prior SOTs.

Results

TEA Results

Based on the key inputs from the cultivation activities noted above that were applied through NREL's biomass production model (i.e., utilizing the SOT productivity, harvest density, and [BAT-modeled] pond evaporation data for the seasonal strain production cases, coupled with the fixed HCSD compositional attributes as discussed above), the resulting MBSP costs are presented in Figure 2. For reference, Figure 2 also shows the estimated SOT costs for an alternative fully lined pond scenario, as well as the final target design case projections for the same HCSD biomass as established in the biomass design report [3] (although now reflecting the target year as 2030 for ultimately achieving 25 g/m²/day annual productivity). All current, back-cast, and future costs reflected here are consistent with the latest financial parameters based on 2016\$ and 21% tax rates, as applied universally for all BETO Platform models. **The resulting MBSP was estimated as \$683/ton AFDW in 2016\$ for the “unlined pond” base case when reflecting ASU evaporation rates/blowdown demands, which would reduce to \$603/ton if instead reflecting FA evaporation rates, as was the basis for the 2015–2016 SOTs—** compared to the 2030 design case target of \$488/ton (again in 2016\$, maintained as the basis for the remainder of this discussion unless otherwise noted). SOT costs for the “fully lined” alternative scenario would increase up to 28% to \$853/ton or \$772/ton for the ASU and FA evaporation basis, respectively. While the cost of fluazinam usage during warm season months is not explicitly included in this *n*th-plant analysis, a sensitivity analysis reflecting the added costs of fluazinam was conducted under related TEA activities for DISCOVER. Based on an approximate average dosing frequency of 1.5 weeks in the spring and fall and 2 weeks in the winter at a 1-ppm dosing concentration for both the UTEX393 and 26B-AM strains, coupled with an assumed fungicide cost of \$10/kg, **including an explicit cost for fluazinam usage would add an estimated \$8/ton to the overall MBSP** (roughly 1% MBSP increase).

The SOT cost values are strongly tied to productivity, estimated at an annual average of 18.4 g/m²/day (AFDW) for the DISCOVER/ASU cultivation activities described above, representing a 16% improvement in annual average cultivation productivity relative to the FY 2019 SOT basis (15.9 g/m²/day) [7]. Notably, this also represents a 2.2-fold improvement relative to the initial 8.5 g/m²/day benchmark in the FY 2015 SOT as noted earlier. While this highlights substantial progress over the past five years, further improvements still remain to achieve the final goal of 25 g/m²/day by 2030, or 20 g/m²/day by 2025 (as a plausible interim case on the path to 2030). Ongoing work under the DISCOVER consortium is aiming to set out-year goals around these parameters in order to keep progress on track over the next several years. Relative to historical progress made to date (productivity improvements of 7%, 13%, 14%, 36%, and 16% in 2016–2020 relative to each preceding year), it is unlikely such substantial improvements will be sustainable on such a level moving forward indefinitely, but fortunately a lower degree of improvement on the order of 3% year-over-year is all that must be demonstrated over the next 10 years in order to ultimately achieve the 2030 goal of 25 g/m²/day. Additionally, a key further improvement moving forward will be to shift composition toward less replete, lower-protein biomass without sacrificing on productivity gains. DISCOVER is prepared to support continued improvements, with internal goals aspiring to exceed such levels over coming years. The data and resulting MBSP values presented here provide value as a public benchmark, as they are based on transparent, long-term growth trials with public documentation of a vast data set.

Additionally, for the recent FY 2017–2020 SOTs based on local evaporation rates pertinent to ASU’s site, salt management/disposal costs were also seen to incur substantial cost penalties relative to those details at a Gulf Coast site with less net evaporation such as FA. As noted above, given significantly higher “net” evaporation rates (inclusive of precipitation considerations) for the ASU site (Phoenix, Arizona) versus the FA site (Vero Beach, Florida), this requires substantially more removal of blowdown as shown in Figure 1 to maintain salt levels within strain tolerance. In turn, the blowdown must be disposed of, assuming costs commensurate with deep-well saline injection. The costs for the injection/disposal step is maintained at \$1.80/m³ (2016\$) [16–20], consistent with the FY 2017 SOT discussion. In addition, as described earlier, two other mitigation measures were also maintained, including (1) evaporation ponds on the blowdown waste stream to reduce overall volumes, and (2) increasing salt tolerance limits up to 50 ppt (utilized for both FA and ASU evaporation cases). Even after these measures, the MBSP contributions attributed to salt disposal are roughly \$22/ton higher for the FY 2020 SOT based on ASU versus FA evaporation. Given that this incurs such a large and artificial penalty on MBSP, to control for this variable and provide a more consistent comparison against the FY 2015–16 SOTs, the alternative FY 2020 SOT scenario based on FA evaporation rates is important to consider given that *overall, this basis reduces MBSP costs by \$80/ton relative to the ASU evaporation basis*. Thus, this reiterates an important conclusion that *arid climates with high evaporation/low precipitation rates (such as the U.S. Southwest) are not ideal locations for siting algal cultivation facilities relative to lower-evaporation locations (such as the U.S. Gulf Coast), if focused on saline cultivation*, due purely to challenges in how to manage salt. While prior harmonization work had assumed this logic (i.e., in placing a high priority on minimizing water losses), the present work quantifies the TEA penalty for this issue. However, given that the primary resources and expertise in algal cultivation to support the SOTs reside at ASU, we will continue to report on SOT MBSPs attributed to both Arizona and Florida evaporation rates, assuming similar performance could be achieved at the latter location (again, BAT modeling may help resolve this in future SOTs).

Figure 2. Biomass production MBSP results and cost breakdowns by major contributions for 2020 SOT, compared against 2015–2019 SOTs and 2025/2030 projections for reference [3] (2016\$, all based on HCSD composition).

First two 2017–2020 SOT bars are based on ASU cultivation performance with ASU local evaporation rates; third bar is based on ASU cultivation performance with Florida Algae (FA) evaporation rates.

Table 5. Technical Overview Table for Cost and Process Metrics Associated with Current and Back-Cast Algal Biomass SOT Cases, Compared Alongside Future 2025–2030 Projections – FA Evaporation Basis (Costs in 2016\$)

Processing Area Cost Contributions & Key Technical Parameters	Metric	2015 SOT ^a	2016 SOT – ATP ^{3 a}	2016 – ABY1 ^a	2017 SOT (FA Evap) ^{a, b}	2018 SOT (FA Evap) ^{a, c}	2019 SOT (FA Evap) ^{a, d}	2020 SOT (FA Evap) ^{a, e}	2025 Projection	2030 Projection
Biomass Selling Price (With Liners)	\$/ton AFDW	\$1,142 (\$1,508)	\$1,089 (\$1,433)	\$960 (\$1,250)	\$909 (\$1,211)	\$824 (\$1,090)	\$670 (\$866)	\$603 (\$772)	\$602	\$488
Production Cost (With Liners)	\$/ton AFDW	\$999 (\$1,365)	\$947 (\$1,291)	\$824 (\$1,115)	\$775 (\$1,078)	\$704 (\$970)	\$556 (\$752)	\$500 (\$669)	\$509	\$400
Harvest/Dewatering Cost	\$/ton AFDW	\$105	\$110	\$107	\$97	\$87	\$82	\$75	\$62	\$63
Other Cost (Facility Circulation, Storage)	\$/ton AFDW	\$38	\$32	\$28	\$36	\$33	\$32	\$28	\$32	\$25
Net Biomass Production Yield ^f	Ton AFDW/ acre-year	12.4	13.2	15.6	15.0	17.0	23.1	26.7	29.9	37.2
Cultivation Productivity (Annual Average)	g/m ² /day	8.5	9.1	10.7	10.3	11.7	15.9	18.4	20	25
Max Seasonal Production Variability	Max:Min Productivity	2.3:1	2.6:1	3.6:1	2.6:1	2.0:1	4.2:1	3.8:1	3:1	3:1
Biomass Harvest Concentration	g/L AFDW	0.27	0.26	~0.5	0.43	0.51	0.49	0.46	0.5	0.5
Total Farm Power Demand	KWh/ton AFDW	860	831	739	717	647	529	486	395	334

^a Base case assumes *n*th-plant facility utilizing low-cost unlined ponds; alternative SOT scenarios consider fully lined ponds with resulting costs shown in parentheses

^b FY 2017 values shown are for FA evaporation basis for consistency with prior FY 2015–2016 SOTs and future projection cases. ASU evaporation basis values are as follows: biomass selling price = \$1,063/ton (\$1,366/ton lined); production cost = \$896/ton (\$1,199/ton lined); harvest/dewatering cost = \$93/ton; other cost = \$74/ton.

^c FY 2018 values shown are for FA evaporation basis for consistency with prior FY 2015–2016 SOTs and future projection cases. ASU evaporation basis values are as follows: biomass selling price = \$955/ton (\$1,222/ton lined); production cost = \$806/ton (\$1,073/ton lined); harvest/dewatering cost = \$84/ton; other cost = \$65/ton.

^d FY 2019 values shown are for FA evaporation basis for consistency with prior FY 2015–2016 SOTs and future projection cases. ASU evaporation basis values are as follows: biomass selling price = \$764/ton (\$961/ton lined); production cost = \$629/ton (\$827/ton lined); harvest/dewatering cost = \$79/ton; other cost = \$55/ton.

^e FY 2020 values shown are for FA evaporation basis for consistency with prior FY 2015–2016 SOTs and future projection cases. ASU evaporation basis values are as follows: biomass selling price = \$683/ton (\$853/ton lined); production cost = \$563/ton (\$733/ton lined); harvest/dewatering cost = \$71/ton; other cost = \$49/ton.

^f Net yield to downstream conversion, after blowdown/short-term storage losses

Sustainability Metric Indicators

In addition to the TEA results noted above, we also report here on associated sustainability “indicators” attributed to the algae farm SOT model. In keeping with recent BETO guidance for all formal life cycle assessment (LCA) sustainability metrics to be handled by Argonne National Laboratory (ANL) to ensure no inconsistencies in such metrics versus NREL-calculated values (i.e., using GREET versus SimaPro), we avoid reporting on LCA parameters such as greenhouse gas emissions or fossil energy consumption in this report (but are currently working to provide the input/output inventories to partners at ANL). Instead, Table 6 summarizes key sustainability indicators as may be taken directly from the Aspen Plus process models. Namely, for the biomass production SOT this includes areal biomass yields, carbon efficiency from delivered CO₂, facility power demand, and water consumption. On the latter parameter, net makeup water demands are listed, but because this SOT and all future projections are to be based on saline cultivation per recent BETO guidance, this does not count against formal consumptive water use, which is based strictly on freshwater consumption (zero in the case of the algal biomass production models). The process input/output inventories furnished to ANL for subsequent LCA supply chain sustainability analysis (SCSA) are summarized in Appendix B.

Table 6. Sustainability Indicators for FY 2020 SOT Biomass Model

Parameter		FY 2020 SOT Evaporation Basis	
		ASU Evap	FA Evap
Net Biomass Yield to Conversion	Ton/acre-yr AFDW ^a	26.7	26.7
Carbon Efficiency to Biomass	% of delivered CO ₂ ^b	90%	90%
Electricity Import	kWh/ton AFDW	591	486
Natural Gas Import	MJ/ton AFDW	N/A	N/A
Water Consumption (SALINE ONLY)	gal/ton AFDW ^c	114,111	8,566
Water Consumption (SALINE ONLY)	m ³ /day ^c	174,500	13,136

^a Net areal biomass yield after accounting for blowdown/short-term storage losses (output to conversion)

^b No SOT data available to date; fixed constant in SOT models at 90% consistent with targets from algae farm design case

^c Values are for saline makeup water only, does not count against formal BETO metrics based on freshwater consumption.

Concluding Remarks

Based on incorporating experimentally observed performance metrics for algal cultivation as achieved under DISCOVER efforts into NREL's latest algal biomass production model (while leaving all other process and costing assumptions for non-cultivation operations unchanged relative to the 2016 biomass design case), **the estimated base case SOT minimum biomass selling price is \$683/ton AFDW in 2016\$ for ASU site evaporation/blowdown rates, or \$603/ton for Florida Algae evaporation/blowdown rates.** This represents the best available seasonal cultivation data attributed to ASU production of *Scenedesmus*, *Monoraphidium*, and *Picochlorum* strains rotated seasonally (overlayed with NREL's HCSD biomass composition), assuming an n^{th} -plant model utilizing low-cost unlined ponds. Alternatively, a scenario employing fully lined ponds would translate to a considerably higher SOT biomass cost of \$853/ton and \$772/ton for the ASU and FA evaporation cases, respectively. The SOT MBSP value is tied primarily to ASU-demonstrated productivity rates, calculated at 18.4 g/m²/day AFDW as seasonal averages for the AzCATI site. This represents a 16% improvement in productivity over the FY 2019 SOT basis at 15.9 g/m²/day, leading to a roughly 10% reduction in SOT biomass cost. This is another considerable improvement, and likely attributed to continued focus on a hypothesis-driven approach to cultivation based on discussions and activities in consultation with the DISCOVER consortium, including the use of high-performing strains and different pond operational strategies (i.e., incorporating fungicide to mitigate contamination during appropriate seasonal periods). Three out of four seasons were shown to result in higher productivities, ranging between 16%–32% improvement over FY 2019 performance levels, including a notable increase over already high summer productivities attributed to the use of a new strain optimally suited to hot summer temperatures. While the cost of fluazinam fungicide utilized experimentally was not explicitly included in this n^{th} -plant analysis, a separate sensitivity case estimates that its usage would incur a trivial penalty of roughly \$8/ton to overall MBSPs.

Given the significant logistical and cost challenges attributed to salt management and disposal in the case of saline cultivation, which are intensified in arid regions with high evaporation (as indicated by the MBSP differences above between ASU and FA evaporation), from strictly a practicality cost minimization standpoint this points to either (1) utilizing freshwater cultivation in those areas (which is also a challenge given limited freshwater resources in those same areas), or (2) siting commercial facilities in low-evaporation regions (e.g., U.S. Gulf Coast area). In light of the artificial cost impact incurred around evaporation/salt blowdown disposal, which is otherwise irrelevant of scientific advancements, a more consistent basis for comparison to prior SOTs may be the ASU data overlaid with FA evaporation rates (\$603/ton MBSP = 10% reduction in MBSP relative to the FY 2019 SOT on this basis). In fact, because FA annual average productivity had originally been seen to be similar or marginally better than at ASU for both the FY 2015 and FY 2016 SOT data sets under ATP³, the MBSP may plausibly be expected to be even lower than the \$603/ton value if the FA site were still available.

As part of future SOT efforts, it is hoped that cultivation trials can credibly demonstrate substantial improvements in both productivity as well as compositional quality, in moving toward 2030 design case targets. On the latter metric, in order to improve compositional quality particularly towards higher-carbohydrate/lower-protein biomass, *Nannochloropsis* has been eliminated from seasonal strain rotations in recent SOTs, recognizing it will never achieve a

suitable composition for CAP conversion particularly around carbohydrates. Additionally, other gaps that could be better addressed in future SOT iterations include tracking (and ultimately improving on) CO₂ utilization efficiency, TEA implications of cultivation dynamics around batch versus semi-continuous harvesting, cost trade-offs between contamination mitigation measures versus crash frequency or growth rate penalties, and experimental demonstrations for dewatering efficacy or at least propensity for a strain to settle. Likewise, a more detailed TEA approach to quantifying economic implications for seasonal strain rotation to weigh penalties versus benefits relative to use of a single strain year-round would be useful moving forward. Such details as these are being considered in greater granularity in support of a pre- n^{th} -plant “operational baseline” metric under the TEA subtask of the DISCOVR consortium, with further work planned in FY 2021.

Consistent with prior SOT conclusions, we reiterate that improving cultivation performance (yields/compositions) and controlling cultivation costs will be key to achieving economically viable algal biofuels for any conversion pathway option. On the cost control side, this would call for eventually demonstrating the viable use of large-scale unlined growth ponds on the order of 10 acres in size [3], or potentially pursuing low-cost PBR/pond hybrid systems as recently published in literature [9]. Additionally, algal wastewater treatment may provide alternative cost benefits including reduced nutrient costs and water treatment credits [21]; this point has been recently reinforced through discussions with wastewater treatment technology providers in industry, who are looking to scale-up algal wastewater treatment in the near term, as well as through internal NREL TEA modeling to quantify economic incentives for algal water treatment scenarios, albeit at more limited national scalability for commodity biomass/biofuel production potential.

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Appendix A. TEA Summary Sheet for Base Case Biomass Cultivation SOT Benchmark Model (FA and ASU Evaporation MBSP Scenarios, 2016 Dollars)

Algal Biomass Production Process Engineering Analysis

2020 SOT with FA Evaporation Rates
All Values in 2016\$

MBSP (Minimum Biomass Selling Price): \$603 /US Dry Ton (AFDW)

Contributions:	CO2 and Nutrients	\$127 /US Dry Ton
	Cultivation	\$361 /US Dry Ton
	Other Production	\$119 /US Dry Ton
Total Biomass Production (AFDW Basis)		0.13 MM US Ton/yr
Total Biomass Yield (AFDW Basis)		26.7 US Ton/acre/yr
		59.9 Metric tonne/ha/yr

Internal Rate of Return (After-Tax)	10%
Equity Percent of Total Investment	40%

Capital Costs	
Production ponds (AVERAGE of 10-acre Cases)	\$151,000,000
Inoculum Ponds	\$15,400,000
CO2 Delivery	\$5,500,000
Makeup Water Delivery + On-Site Circulation	\$9,000,000
Dewatering	\$37,900,000
Storage	\$3,600,000
Total Installed Equipment Cost	\$222,400,000
Added Direct + Indirect Costs (% of TCI)	\$143,700,000 39%
Total Capital Investment (TCI)	\$366,100,000
Installed Equipment Cost/Annual US dry ton biomass	\$1,663
Total Capital Investment/Annual US dry ton biomass	\$2,738
Loan Rate	8.0%
Term (years)	10
Capital Charge Factor (Computed)	0.127

Manufacturing Costs (\$/US Ton AFDW Biomass product)	
CO2	\$98.84
Nutrients	\$23.66
Salt Disposal	\$4.87
Power	\$33.12
Other Operating Costs	\$5.39
Fixed Costs	\$87.66
Capital Depreciation	\$81.52
Average Income Tax	\$28.00
Average Return on Investment	\$239.46

Manufacturing Costs (\$/yr)	
CO2	\$13,200,000
Ammonia	\$2,300,000
Diammonium Phosphate	\$900,000
Power	\$4,400,000
Chilled Water Utility	\$700,000
Fixed Costs	\$11,700,000
Capital Depreciation	\$10,900,000
Average Income Tax	\$3,700,000
Average Return on Investment	\$32,000,000

Algal Biomass Production Process Engineering Analysis

2020 SOT with ASU Evaporation Rates
All Values in 2016\$

MBSP (Minimum Biomass Selling Price): \$683 /US Dry Ton (AFDW)

Contributions:	CO2 and Nutrients	\$150 /US Dry Ton
	Cultivation	\$424 /US Dry Ton
	Other Production	\$136 /US Dry Ton
Total Biomass Production (AFDW Basis)		0.13 MM US Ton/yr
Total Biomass Yield (AFDW Basis)		26.7 US Ton/acre/yr 59.8 Metric tonne/ha/yr
Internal Rate of Return (After-Tax)		10%
Equity Percent of Total Investment		40%

Capital Costs	
Production ponds (AVERAGE of 10-acre Cases)	\$175,300,000
Inoculum Ponds	\$18,000,000
CO2 Delivery	\$5,500,000
Makeup Water Delivery + On-Site Circulation	\$10,000,000
Dewatering	\$37,600,000
Storage	\$9,900,000
Total Installed Equipment Cost	\$256,300,000
Added Direct + Indirect Costs (% of TCI)	\$158,300,000 38%
Total Capital Investment (TCI)	\$414,600,000
Installed Equipment Cost/Annual US dry ton biomass	\$1,922
Total Capital Investment/Annual US dry ton biomass	\$3,110
Loan Rate	8.0%
Term (years)	10
Capital Charge Factor (Computed)	0.127

Manufacturing Costs (\$/US Ton AFDW Biomass product)	
CO2	\$99.11
Nutrients	\$23.87
Salt Disposal	\$26.91
Power	\$40.29
Other Operating Costs	\$5.41
Fixed Costs	\$91.23
Capital Depreciation	\$93.01
Average Income Tax	\$31.64
Average Return on Investment	\$271.48

Manufacturing Costs (\$/yr)	
CO2	\$13,200,000
Ammonia	\$2,300,000
Diammonium Phosphate	\$900,000
Power	\$5,400,000
Chilled Water Utility	\$700,000
Fixed Costs	\$12,200,000
Capital Depreciation	\$12,400,000
Average Income Tax	\$4,200,000
Average Return on Investment	\$36,200,000

Appendix B. Life Cycle Inventory for 2020 SOT Algae Farm Model

SOT Front-End Input and Output Data for the Modeled Algae Production Facility (10-Acre Average Base Case). (Note: Daily rates shown below are based on annual averages over all modeled seasons based on 24-hour day)

Products, kg/hr	Annual Average Rates FA Evap.	Annual Average Rates ASU Evap.
Algal biomass (AFDW) ^a	15,312	15,269
Algal biomass (total including ash) ^a	15,692	15,648
Resource Consumption, kg/hr		
CO ₂ (counted as biogenic)	34,030	34,030
Ammonia	306	308
DAP	148	149
Total process water input (<i>SALINE</i>) ^b	547,338	7,270,840
Electricity demand, kW	8,030	9,740
Output Streams, kg/hr		
Water in biomass product stream	61,867	61,693
Water lost to blowdown	91,091	1,002,051
Algae lost in blowdown	5	48
Air Emissions, kg/hr		
Water lost to evaporation	379,306	6,140,542
CO ₂ outgassing from ponds (counted as biogenic)	3,403	3,403
O ₂ to atmosphere	26,399	26,399

^a Total after 1% algae loss for storage

^b Total water input, including the amount contained in the biomass product stream sent to conversion (in many cases, a large fraction of this water is ultimately recycled back to ponds from downstream conversion steps); all makeup water is saline.